## Appendix A. Data Management Rules

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## Acronyms

| cPAH | carcinogenic polycyclic aromatic hydrocarbon |
| :--- | :--- |
| DDD | dichlorodiphenyldichloroethane |
| DDE | dichlorodiphenyldichloroethylene |
| DDT | dichlorodiphenyltrichloroethane |
| EPA | US Environmental Protection A gency |
| HPAH | high-molecular-weight polycyclic aromatic hydrocarbon |
| HpCDD | heptachlorodibenzo-p-dioxin |
| HpCDF | heptachlorodibenzofuran |
| HxCDD | hexachlorodibenzo-p-dioxin |
| HxCDF | hexachlorodibenzofuran |
| LPAH | low-molecular-weight polycyclic aromatic hydrocarbon |
| PAH | polycyclic aromatic hydrocarbon |
| PCB | polychlorinated biphenyl |
| PeCDD | pentachlorodibenzo-p-dioxin |
| PeCDF | pentachlorodibenzofuran |
| PEF | potency equivalency factor |
| QC | quality control |
| RIIFS | remedial investigation and feasibility study |
| RL | reporting limit |
| SIM | selected ion monitoring |
| SMS | Washington State Sediment M anagement Standards |
| SVOC | semivolatile organic compound |
| TCDD | tetrachlorodibenzo-p-dioxin |
| TCDF | tetrachlorodibenzofuran |
| TEF | toxic equivalency factor |
| TEQ | toxic equivalent |
| WHO | World Health Organization |

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## A. 1 Data Management Rules

Data management rules being followed for this data compilation are the same as those applied to the remedial investigation/ feasibility study (RI/ FS) dataset, except as noted in this section. Rules summarized in this appendix include those used for averaging duplicate or replicate samples (Section A.1), selecting the preferred result if more than one result was reported for a chemical (Section A.2), handling significant figures and rounding (Section A.3), calculating totals when results are summed for individual components (Section A.4), calculating toxic equivalents (TEQs) for polychlorinated biphenyl (PCB) congeners and dioxin/ furan congeners (Sections A. 5 and A.6), and calculating carcinogenic polycyclic aromatic hydrocarbons (cPAHs) (Section A.7). Final data rules are provided in A ppendix C of the Pre-Design Studies Work Plan (Windward and Integral [in prep]).

## A.1.1 Averaging Laboratory Duplicate or Replicate Samples

Contaminant concentrations obtained from the analysis of laboratory duplicates or replicates (i.e., two or more analyses on the same sample) were averaged for a closer representation of the "true" concentration than that provided by the results of a single analysis. Averaging rules were dependent on whether the individual results were detected concentrations or reporting limits (RLs) for non-detected analytes. If all concentrations were detected for a given parameter, the values were simply averaged arithmetically. If all concentrations were non-detected for a given parameter, the minimum RL was reported. If the concentrations were a mixture of detected concentrations and RLs, any two or more detected concentrations were averaged arithmetically, and RLs were ignored. If there was one detected concentration and one or more RLs, the detected concentration was reported. The latter two rules were applied regardless of whether the RLs were higher or lower than the detected concentration.

## A.1.2 Selection of Preferred Results

In some instances, the laboratory generated more than one result for a chemical for a given sample. M ultiple results occurred for several reasons, including:
u The original result did not meet the laboratory's internal quality control (QC) guidelines, and a reanalysis was performed.
$u \quad$ The original result did not meet other project data quality objectives, such as a sufficiently low RL, and a reanalysis was performed.
u Two different analytical methods were used for that chemical.

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In each case, a single result was selected for use. The procedures for selecting the preferred result differed depending on whether a single or multiple analytical methods were used for that chemical.

For the same analytical method, the results were selected using the following guidance:
u If the results were detected and not qualified, then the result from the lowest dilution was selected, unless multiple results from the same dilution were available, in which case the result with the highest concentration was selected.
u If the results were combination of estimated and unqualified detected results, then the unqualified result was selected. This situation most commonly occurred when the original result was outside of the calibration range, thus requiring a dilution. The diluted result within the calibration range was preferentially selected.
u If the results were all estimated, then the result was selected using best professional judgment and considering the rationale for qualification. For example, a result qual ified based on laboratory replicate results outside of QC objectives for precision was preferred to a qualified result that was outside the calibration range.
u If the results were a combination of detected and non-detected results, then the detected result was selected. If there were more than one detected result, the applicable rules for multiple results (as discussed above) were followed.
u If the results were all non-detected, then the lowest RL was selected.
For different analytical methods (i.e., when a specific chemical was analyzed in the same sample using different methods), the following rules were applied:
u For results analyzed using the semivolatile organic compound (SVOC) full-scan (EPA 8270) and selected ion monitoring (SIM) (EPA 8270-SIM) methods, the SIM results were selected.
u For results analyzed using US Environmental Protection A gency (EPA ) M ethod 8081A and any 8270 method (i.e., hexachlorobenzene and hexachlorocyclopentadiene), the 8081A result were selected.

The RI/ FS database rules for the selection of preferred results between two methods (as described above) were revised for the compilation of the pre-design data. In the RI/ FS, the preferred result was selected based on a comparison among the methods of the detection status, RL, and data qualifiers. The revised rules selected the preferred result based on a preference for method.

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## A.1.3 Significant Figures and Rounding

The analytical laboratories reported results with various numbers of significant figures depending on the instrument, parameter, and concentration relative to the RL. The reported (or assessed) precision of each observation was explicitly stored in the project database as a record of the number of significant figures assigned by the laboratory. The tracking of significant figures became important when calculating averages and performing other data summaries.
When a calculation involved addition, such as totaling PCBs or polycyclic aromatic hydrocarbons (PAHs), the calculation was only as precise as the least precise number that went into the calculation. For example (assuming two significant figures):
$210+19=229$ was reported as 230 because 19 was only reported to 2 significant digits, and the enhanced precision of the trailing 0 in the number 210 was not significant.
When a calculation involved multiplication or division, such as carbon normalization, the original figures for each value were carried through the calculation (i.e., individual values were not adjusted to a standard number of significant figures; instead, the appropriate adjustment was made to the resultant value at the end of the calculation). The result was rounded at the end of the calculation to reflect the value with the fewest significant figures used in the calculation. For example:
$59.9 \times 1.2=71.88$ was reported as 72 because there were 2 significant figures in the number 1.2.
When rounding, if the number following the last significant figure was less than 5 , the digit was left unchanged. If the number following the last significant figure was equal to or greater than 5 , the digit was increased by 1 .

## A.1.4 Calculating Totals

Total PCBs, total dichlorodiphenyltrichloroethanes (DDTs), total PAHs, total chlordane, total xylenes, and total nitrosamines were calculated by summing the detected values for the individual components (e.g., A roclor mixtures or individual congeners for total PCBs). For samples in which none of the individual components were detected, the total value was given as the highest RL of any individual component, and assigned a U-qualifier (no detected concentrations). No sum was cal culated in cases where $50 \%$ or less of the components were analyzed.
Concentrations for anal yte sums were cal culated using the following components:
u Total PCBs were calculated, in accordance with the methods of the Washington State Sediment M anagement Standards (SMS), using only detected values for all Aroclor mixtures. For individual samples in which none of the Aroclor mixtures were detected, total PCBs were given a value equal to the highest RL of the Aroclors and assigned a U-qualifier (no detected concentrations).
u Total low-molecular-weight PAHs (LPAHs), high-molecular-weight PAHs (HPAHs), PAHs, and benzofluoranthenes were also calculated in accordance with the methods of the SMS. Total LPA Hs were the sum of detected concentrations for naphthalene, acenaphthylene, acenaphthene, fluorene, phenanthrene, and anthracene. Total HPA Hs were the sum of detected concentrations for fluoranthene, pyrene, benzo(a)anthracene, chrysene, total benzofluoranthenes, benzo(a)pyrene, indeno(1,2,3,-c,d)pyrene, dibenzo( $\mathrm{a}, \mathrm{h}$ )anthracene, and benzo( $\mathrm{g}, \mathrm{h}, \mathrm{i}$ )perylene. Total benzofluoranthenes were the sum of the b (i.e., benzo(b)fluoranthene), j, and k isomers.

Because the j isomer is rarely quantified, the total benzofluoranthenes sum was typically calculated with only the $b$ and $k$ isomers. In cases where the laboratory provided total benzofluoranthenes instead of or in addition to the b and k isomers, the laboratory result was reported, and no sum was calculated. For samples in which all individual compounds within any of the three groups described above were non-detected, the highest RL for that sample represented the sum.
u Total DDTs were calculated using only detected values for the DDT isomers: 2,4'-dichlorodiphenyldichloroethane (DDD); 4,4'-DDD;
2,4'-dichlorodiphenyldichloroethylene (DDE); 4,4'-DDE; 2,4'-DDT; and 4,4'-DDT. For individual samples in which none of the isomers were detected, total DDTs were given a value equal to the highest RL of the six isomers and assigned a U-qualifier (no detected concentrations).
u Total chlordane was calculated using only detected values for the following compounds: alpha-chlordane, gamma-chlordane, oxychlordane, cis-nonachlor, and trans-nonachlor. For individual samples in which none of these compounds were detected, total chlordane was given a value equal to the highest RL of the five compounds listed and assigned a U-qualifier (no detected concentrations).
u Total xylene was calculated using only detected values for m,p-xylene and o-xylene. For individual samples in which neither of these compounds were detected, total xylene was given a value equal to the higher RL of the two compounds listed and assigned a U-qualifier (no detected concentrations).
u Total nitrosamines were calculated using only detected values for n-nitrodiethylamine, n-nitrosodimethylamine, n-nitroso-di-n-butylamine, n-nitroso-di-n-propylamine, and n-nitrosodiphenylamine. For individual samples in which none of these compounds were detected, total nitrosamines were given a value equal to the highest RL of the five compounds listed and assigned a U-qualifier (no detected concentrations).

## A.1.5 Calculation of PCB Congener Toxic Equivalents

PCB congener TEQs were calculated using the World Health Organization (WHO) consensus toxic equivalency factor (TEF) values for mammals (Van den Berg et al. 1998; Van den Berg et al. 2006), as presented in Table 1. The TEQ was calculated as the sum of each PCB congener concentration multiplied by the corresponding TEF value. When the PCB congener concentration was reported as non-detected, then the TEF was multiplied by onehalf the RL.

Table A-1.PCB congener TEF values

| PCB Congener <br> No. | TEF Value for Mammals <br> (unitless) $^{\mathbf{a}}$ |
| :--- | :---: |
| 77 | 0.0001 |
| 81 | 0.0003 |
| 105 | 0.00003 |
| 114 | 0.00003 |
| 118 | 0.00003 |
| 123 | 0.00003 |
| 126 | 0.1 |
| 156 | 0.00003 |
| 157 | 0.00003 |
| 167 | 0.00003 |
| 169 | 0.03 |
| 189 | 0.00003 |

a From Van den Berg et al. (2006).
PCB - polychlorinated biphenyl
TEF - toxic equivalency factor

## A.1.6 Calculation of Dioxin/Furan Congener TEQs

Dioxin/ furan congener TEQs were calculated using the WHO consensus TEF values for mammals (Van den Berg et al. 1998; Van den Berg et al. 2006), as presented in Table 2. The TEQ was calculated as the sum of each dioxin/ furan congener concentration multiplied by the corresponding TEF value. When the dioxin/ furan congener concentration was reported as non-detected, then the TEF was multiplied by one-half the RL.

Table A-2. Dioxin/furan congener TEF values

| Dioxin/Furan Congener | TEF Value for <br> Mammals <br> (unitless) ${ }^{\text {a }}$ |
| :--- | :---: |
| $1,2,3,4,6,7,8-$ HpCDF | 0.01 |
| $1,2,3,4,6,7,8-H p C D D$ | 0.01 |
| $1,2,3,4,7,8,9-H p C D F$ | 0.01 |

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| Dioxin/Furan Congener | TEF Value for <br> Mammals <br> (unitless) $^{\text {a }}$ |
| :--- | :---: |
| $1,2,3,4,7,8-\mathrm{HxCDF}$ | 0.1 |
| $1,2,3,4,7,8-\mathrm{HxCDD}$ | 0.1 |
| $1,2,3,6,7,8-\mathrm{HxCDF}$ | 0.1 |
| $1,2,3,6,7,8-\mathrm{HxCDD}$ | 0.1 |
| $1,2,3,7,8,9-\mathrm{HxCDF}$ | 0.1 |
| $1,2,3,7,8,9-\mathrm{HxCDD}$ | 0.1 |
| $1,2,3,7,8-\mathrm{PeCDF}$ | 0.03 |
| $1,2,3,7,8-\mathrm{PeCDD}$ | 1 |
| $2,3,4,6,7,8-\mathrm{HxCDF}$ | 0.1 |
| $2,3,4,7,8-\mathrm{PeCDF}$ | 0.3 |
| $2,3,7,8-\mathrm{TCDF}$ | 0.1 |
| $2,3,7,8-\mathrm{TCDD}$ | 1 |
| OCDF | 0.0003 |
| OCDD | 0.0003 |

a From Van den Berg et al. (2006).
HpCDD - heptachlorodibenzo-p-dioxin PeCDD - pentachlorodibenzo-p-dioxin
HpCDF - heptachlorodibenzofuran
HxCDD - hexachlorodibenzo-p-dioxin
PeCDF - pentachlorodibenzofuran
HxCDF - hexachlorodibenzofuran
TCDD - tetrachlorodibenzo-p-dioxin

OCDD - octachlorodibenzo-p-dioxin
TCDF - tetrachlorodibenzofuran

OCDF - octachlorodibenzofuran


## A.1.7 Calculation of Carcinogenic Polycyclic Aromatic Hydrocarbons

cPAH values were calculated using potency equivalency factor (PEF) values (California EPA 2009) based on the individual PAH component's relativetoxicity to benzo(a)pyrene. PEF values are presented in Table 3. The cPAH value was cal culated as the sum of each individual PAH concentration multiplied by the corresponding PEF value. When the individual PAH component concentration was reported as non-detected, then the PEF was multiplied by onehalf the RL.

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Table A-3.cPAH PEF values

| cPAH | PEF Value <br> (unitless) |
| :--- | :---: |
| Benzo(a)pyrene | 1 |
| Benzo(a)anthracene | 0.1 |
| Benzo(b)fluoranthene | 0.1 |
| Benzo(k)fluoranthene | 0.1 |
| Chrysene | 0.01 |
| Dibenz(a,h)anthracene | 0.4 |
| Indeno(1,2,3-cd)pyrene | 0.1 |

a PEFs for CPAHs are defined by California EPA (2009) by dividing the inhalation unit risk factor for the compound by the inhalation unit risk factor for benzo[a]pyrene.
cPAH - carcinogenic polycyclic aromatic hydrocarbon
EPA - US Environmental Protection Agency
PEF - potency equivalency factor

## A. 2 References

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